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1 **Distribution and trajectories of floating and benthic marine macrolitter in**
2 **the south-eastern North Sea**

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17

18 **Abstract**

19 In coastal waters the identification of sources, trajectories and deposition sites of marine litter
20 is often hampered by the complex oceanography of shallow shelf seas. We conducted a multi-
21 annual survey on litter at the sea surface and on the seafloor in the south-eastern North Sea.
22 Bottom trawling was identified as a major source of marine litter. Oceanographic modelling
23 revealed that the distribution of floating litter in the North Sea is largely determined by the
24 site of origin of floating objects whereas the trajectories are strongly influenced by wind drag.
25 Methods adopted from species distribution modelling indicated that resuspension of benthic
26 litter and near-bottom transport processes strongly influence the distribution of litter on the
27 seafloor. Major sink regions for floating marine litter were identified at the west coast of
28 Denmark and in the Skagerrak. Our results may support the development of strategies to
29 reduce the pollution of the North Sea.

30

31 **Keywords** North Sea, marine litter, seafloor, surface, composition, trajectory

32

33 **Introduction**

34 Marine litter has accumulated over recent decades in all habitats of the world's ocean
35 from urban coastal waters (Browne et al. 2011) to most remote, supposedly unspoiled
36 environments, such as the polar regions (Obbard et al. 2014) and the deep sea (Bergmann &
37 Klages 2012, Pham et al. 2014). Marine litter is causing harm to a great variety of organisms
38 (Kühn et al. 2015), affects the functioning of marine ecosystems (Green et al. 2015) and has
39 adverse economic and human health effects (Thompson et al. 2009). Although numerous
40 national and international legislative measures have been brought into action to combat
41 marine litter (Hastings & Potts 2013) quantities of marine litter are still increasing worldwide
42 (Jambeck et al. 2015) indicating a poor control of major litter sources and inappropriate waste
43 management.

44 There is a consensus that preventing litter from entering the marine environment
45 should be given priority over the removal of litter which has already escaped from controlled
46 waste streams (van Franeker & Law 2015) because the broad and unselective removal of large
47 quantities of litter from the oceans is inevitably associated with a simultaneous extraction of
48 considerable amounts of marine biological production and damage to sensitive environments.
49 However, the ecological impacts of clean-ups can be reduced if extraction activities are
50 focused on strategically selected and easily accessible sites where litter accumulates, for
51 example, due to local oceanographic conditions (Sherman & van Sebille 2016). Therefore, a
52 sustainable and environmentally sound strategy to reduce marine litter requires the
53 identification of major input pathways of marine litter as well as an understanding of the
54 trajectories and deposition sites of litter items in the marine environment.

55 Major sources of marine litter can often be inferred from the litter composition. For
56 example, derelict fishing gear and shipping equipment clearly hint at maritime activities as a
57 source. A spatial correlation of the quantities of certain litter items and their supposed source
58 activity strongly indicate local input and deposition. For example, Tekman et al. (2017)
59 identified intensified fisheries activities around the Svalbard archipelago as a source of
60 increasing amounts of litter on the seafloor of the nearby Fram Strait. Alternatively, a
61 diffuse distribution of litter unrelated to possible source activities may indicate near-bottom
62 transport of objects after deposition. Resuspension and near-bottom transport of seafloor litter

63 may be induced by the joint action of waves and tidal currents as well as other environmental
64 factors, such as seafloor topography and river plumes (Schlining et al. 2013).

65 The global distribution of marine litter and major accumulation zones in the oceans,
66 such as the oceanic subtropical convergences, are well known from extensive field surveys
67 and oceanographic modelling (Lebreton et al. 2012, Maximenko et al. 2012, Cózar et al.
68 2014, Eriksen et al. 2014). However, in coastal waters and shelf sea regions the distribution
69 and trajectories of marine litter are more variable and often less understood due to the diverse
70 coastal sources of marine litter and the difficulties in modelling the trajectories of flotsam in
71 oceanographically complex coastal waters. The prediction of trajectories and distribution
72 patterns of marine litter is further complicated by the poor skill of predicting the sinking of
73 items. The loss of buoyancy of a floating litter item can be induced by colonization by marine
74 organisms (Ye & Andrady 1991) or by water intrusion upon damage, e.g. in PET drinking
75 bottles. Accordingly, the distribution and composition of litter on the seafloor are the result of
76 at-site deposition from local activities, the import of sinking items from the sea surface, and
77 near-bottom transport. The relative importance of the import of sinking items from the surface
78 may become evident from the degree of similarity in the composition and distribution of litter
79 at the sea surface and on the seafloor with strong correlations indicating frequent import from
80 the sea surface whereas no correlation would indicate that other processes than import from
81 the surface shape the distribution and the site specific composition of benthic litter.

82 In this study we investigate the distribution and composition of marine debris at the
83 sea surface and on the seafloor of the south-eastern North Sea. We conducted a multi-annual
84 survey which produced an extensive data base on the spatio-temporal distribution of marine
85 litter with considerable spatial coverage. Oceanographic models were used to investigate
86 forward and backward trajectories of real floating items on the scale of the entire North Sea in
87 order to identify potential source regions and important sites of deposition. Additionally,
88 advanced methods of species distribution modelling were applied to benthic litter to infer the
89 processes that shape the distribution of marine debris on the seafloor.

90

91 **Material and Methods**

92 *Study area*

93 The North Sea, which is part of the north-eastern Atlantic shelf, is a marginal sea with
94 a surface area of 575,300 km² (ICES 1983). The largest exchange of waters between the
95 North Sea and Atlantic Ocean is between the Shetland Islands and Norway. According to Otto

96 et al. (1990) two main water bodies can be distinguished in the North Sea. The water masses
97 in the northern and central parts bear characteristics of oceanic water with surface salinities
98 above 34 (Weichart 1986, Huthnance 1991) and seasonal stratification (Pingree et al. 1978).
99 The southern North Sea receives oceanic waters mainly through the British Channel. This
100 area is subject to large continental runoff resulting in salinities below 30 in the coastal
101 regions. Due to strong tides the shallow southern North Sea waters are permanently well
102 mixed. Tidal and wind forcing in concert drive an anti-clockwise residual circulation along
103 the coasts (Otto et al. 1990, Huthnance 1991, Pohlmann 2006). Several frontal systems
104 separate the oceanic-like water masses of the central North Sea from the coastal waters.

105 The German Bight in the south-eastern North Sea is bordered by the Frisian Islands.
106 Large tidal flats extend between these barrier islands and the coast constituting the major part
107 of the Wadden Sea. This shallow water body is one of the most prominent regions of
108 freshwater influence world-wide being under riverine influence of the rivers Rhine, Meuse,
109 Ems, Weser, Elbe and Eider. The runoff from some of these rivers is considerable (e.g. ~ 1000
110 $\text{m}^3 \text{s}^{-1}$ for the River Elbe; Dippner 1993). Salinity increases from about 30 in the Wadden Sea
111 to 31-33 at the island of Helgoland, which is located about 50 km offshore. Water moves from
112 the German Bight along the Danish coast into the Skagerrak. A pronounced frontal system
113 along the 30 m depth contour separates the coastal waters from the more saline offshore
114 waters of the German Bight (Krause et al. 1986, Budéus 1989, Becker et al. 1992, Dippner
115 1993, Skov & Prins 2001).

116

117 *Quantities and distribution of floating litter*

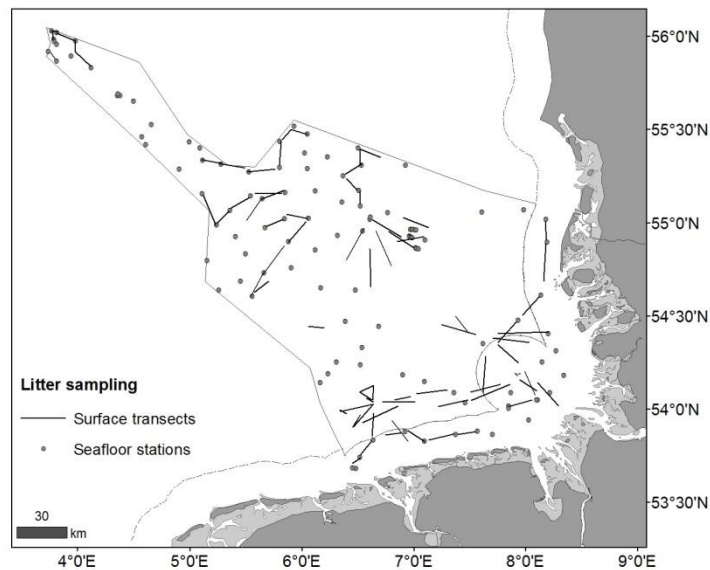
118 Floating marine litter was quantified between 2006 and 2016 on 9 cruises of the
119 German research vessel *Heincke* in the German sector of the North Sea, which were part of
120 extensive marine benthos surveys. Data collected between 2006 and 2008 have been
121 published previously by Thiel et al. (2011). Floating litter items were counted from aboard the
122 vessel during transits between benthos sampling stations. We recorded type and position of all
123 floating macrolitter items (i.e. objects detectable with the naked eye) within a range of 20 to
124 70 m perpendicular to the ship's track for the years 2006 to 2008. During subsequent surveys
125 (years 2014 to 2016) objects were recorded within a shorter range of 10 to 20 m. For each
126 transect we recorded the start and end position from a hand-held GPS. The position of each
127 floating litter item was recorded as the observer position at the time when the item was
128 passing by.

129 Floating litter was recorded only at daytime and during periods of good visibility.
130 During the observations the ship speed varied between about 5 and 11 knots. In total, 78 sea
131 surface transects were surveyed (Figure 1). The length of transects varied between 4.7 and
132 25.1 km. The strip transect method was used to calculate the density D (items km^{-2}) of
133 floating litter items for each transect (for details see Hinojosa et al. 2011) using the following
134 equation:

$$135 \quad D = N / ((W / 1000) \cdot L) \quad \text{Eq. 1}$$

136 where N = the number of items counted, W = the width of the transect and L = the length of
137 the transect in km.

138 Given the patchy distribution of flotsam visual ship based observations along extended
139 transects are a particularly good method to determine the quantity and distribution of surface
140 litter. The exact position of each single item is recorded allowing for a precise description of
141 mesoscale variations in litter densities. Additionally, the considerable area covered by single
142 transects allows for integration at different spatial scales to obtain representative average
143 densities.



144

145 **Figure 1** Positions of surface transects (black lines) and bottom trawls (black dots) for the
146 quantification of floating and benthic marine litter in the south-eastern North Sea. The solid
147 grey line depicts the German Exclusive Economic Zone; the dashed line depicts national
148 territorial waters.

149

150 *Quantities and distribution of litter on the seafloor*

151 Densities of litter on the seafloor were determined from beam-trawl hauls which were
152 taken routinely during benthos surveys in the years 2014 and 2016. A 2 m-beam trawl (mesh
153 size: 1 cm) was towed at an average speed of about 1-3 knots. Trawling distance was
154 calculated from the GPS positions where the winch stopped veering (start trawling) and
155 started hauling (end trawling). The beam trawl was operated during day and night time and
156 was largely independent of the sea state allowing for a total of 122 hauls (Figure 1). All litter
157 items retrieved from the net were identified and counted. Photographs were taken of each
158 item. At two stations numerous fibres from fishing nets were tightly knotted to bunches and
159 could not be counted individually. The number of single fibres in those bunches were
160 estimated from photographs. Equation 1 was also used to calculate the density D of benthic
161 litter items (items km^{-2}) for each haul with W = the width of the beam trawl (here: 2 m) and L
162 = the trawling distance in km.

163 For the qualitative description of the litter composition at the sea surface and at the
164 seafloor we adopted the litter categories defined by Thiel (2013): plastics, styrofoam, glass,
165 metal, wood, paper and other. Additionally, we added the category “organic matter” (e.g.
166 peels of citrus fruits). Wooden items were recorded as litter if they showed clear signs of
167 manufacturing. We contrasted the composition of the total surface litter and the total seafloor
168 litter (i.e. data from all years pooled) with a χ^2 -test based on a 2×8 contingency table.

169 We tested for correlation of litter densities at the surface and at the seafloor. Densities
170 of floating litter at the surface were estimated on transects between the seafloor stations. To
171 obtain spatially corresponding density estimates from the surface and the seafloor a 10×10
172 km grid (European Environment Agency reference grid; [http://www.eea.europa.eu/data-and-](http://www.eea.europa.eu/data-and-maps/data/eea-reference-grids-2)
173 [maps/data/eea-reference-grids-2](http://www.eea.europa.eu/data-and-maps/data/eea-reference-grids-2)) was laid over the study area using ArcGIS 10.3.1
174 (www.esri.com/arcgis). For each surface transect and bottom trawl the central position was
175 used as the geographic reference position. Density estimates from the two habitats within the
176 same cell were accepted as spatially corresponding irrespective of the sampling time. If a cell
177 contained more than one count from the same habitat we used the mean of the multiple
178 estimates resulting in a total of 18 corresponding density estimates from the surface and the
179 seafloor. Since a Kolmogorov-Smirnov test confirmed that the data did not follow a Gaussian
180 distribution, a non-parametric Spearman rank correlation was used to test for correlation
181 between surface and seafloor densities.

182

183 *Drift trajectories of floating litter*

184 For each single litter item that was encountered floating in the German North Sea
 185 during the multi-annual survey we modelled the drift trajectory of a Lagrangian particle
 186 forward and backward in time to identify areas of deposition and origin, respectively. We
 187 assumed that surface transport of floating litter items is driven by the joint action of currents
 188 and wind. Trajectories were simulated with a Lagrangian model using the passive tracer
 189 module of the open-source software OpenDrift (Dagestad et al. 2017).

190 Data on current velocities were obtained from the Forecasting Ocean Assimilation
 191 Model Atlantic Margin Model 7 km (FOAM AMM7) configuration of the Copernicus Marine
 192 Environment Monitoring Service (<http://marine.copernicus.eu/>) based on the ocean model
 193 Nucleus for European Modelling of the Ocean (NEMO). The AMM7 setup has a horizontal
 194 resolution of $1/9^\circ$ longitude \times $1/15^\circ$ latitude (ca. 7.4 km) with 51 σ -layers providing daily
 195 mean surface velocity fields. Details of NEMO are given in Madec (2008) and an extensive
 196 validation is provided in O’Dea et al. (2012). The AMM7 model domain covers the North Sea
 197 and the shelf edge. For this study the area was restricted to 48.0 to 61.4°N and -5.0 to 13.0°E
 198 which covers the North Sea, Skagerrak, Kattegat, The British Channel and is cut off north of
 199 the Shetland Islands (Figure S1).

200 Wind data were provided by the German Weather Service (DWD) with a horizontal
 201 resolution of $1/16^\circ$ (ca. 7 km). Due to limited data availability the wind data were divided into
 202 two periods. For the first 2 years (starting on 01/06/2006) the temporal resolution is 3 h and
 203 for the last 8 years and 9 months (ending on 11/3/2017) the temporal resolution is 1 h. All
 204 other properties are identical among the two periods.

205 Marine litter items appear in various forms and sizes and it is impossible to actually
 206 define and specify suitable wind drag coefficients and leeway drift (or deflection angles) as it
 207 is common in search-and-rescue models for specific object classes like ship containers, life
 208 rafts or people with life jackets in the water (see e.g. Breivik et al. 2011, 2013). Therefore,
 209 particle trajectories were simulated using three different hypothetical wind drag coefficients:
 210 0.0%, 0.5% and 1.0%. The wind drag coefficients were multiplied by the wind speed and
 211 added to the current velocity. We used a 2nd-order Runge-Kutta algorithm as advection
 212 scheme. The trajectory of each particle was calculated with the following equations:

213

$$214 \quad dx(t) = \quad - \quad - \quad - \quad \text{Eq. 2}$$

$$215 \quad - \quad - \quad - \quad \text{Eq. 3}$$

216

217 where $dx(t)$, $dy(t)$ are the latitudinal and longitudinal particle displacements, u , v are the
218 velocities at the particle position at $t+\frac{dt}{2}$, u_{wind} and v_{wind} are the wind velocities, w_d is the wind
219 drag coefficient, dt is the time step and u' and v' are random velocities to parametrize
220 subscale turbulent processes. The turbulent diffusion was implemented by adding random
221 walk displacements (Callies et al. 2011; Neumann et al. 2014). For every time step the
222 turbulent velocities are calculated as

223

$$224 \quad u'(t) = \sqrt{\frac{2D}{dt}} R_x(t) \quad \text{Eq. 4}$$

$$225 \quad v'(t) = \sqrt{\frac{2D}{dt}} R_y(t) \quad \text{Eq. 5}$$

226

227 where D is the eddy diffusivity calculated as proposed by Schönfeld et al. (1995) and is 15.9
228 $\text{m}^2 \text{s}^{-1}$ in this setup. R_x and R_y are independent random numbers from a normal distribution
229 with zero mean and variance one. Further details can be found in Callies et al. (2011). The
230 model time step was $dt = 1$ h and the output time step was 1 d.

231 The actual processes of beaching of marine litter are not well understood and,
232 therefore, currently only numerical artefacts of the advection process represent the beaching
233 of particles. When the combined transport vector (ocean current plus wind drag effect) acts on
234 a particle close to shore the particle is pushed over the model boundary onto land (given a
235 long enough time step dt in eqs. 2+3). All particles that so reached land in this study, were
236 considered beached and were not further advected. The time and position of beaching was
237 recorded. Similarly, if a particle exited the domain through an open boundary, e.g. towards the
238 north into the Atlantic, the time and position was saved. For each time step the fraction of
239 particles in each region (Figure S1) is given as percentage of all particles.

240 Observations of a total of 772 floating litter items were considered for the simulation.
241 The litter observation dates were used as the seeding dates of the particles in OpenDrift. The
242 forward simulations started on 25/07/2006 12:00 h (date of the first litter observation) and
243 ended on 11/03/2017 00:00 h, which was the last available velocity data at the time when the
244 simulation was run. Alternatively, the simulation was stopped when all particles were either
245 beached or had left the domain. The latest litter observation was on 10/04/2016 12:00 h.

246 Backward simulations started with the latest observed litter item and ended at the
247 earliest on 01/06/2006 or when all particles were beached. As a consequence of the
248 distribution of the surface transects two clusters of floating particles became apparent: an
249 offshore cluster and a nearshore cluster. The two clusters were separated by the major salinity
250 front of the south-eastern North Sea (Figure S1). Since fronts can efficiently separate water
251 masses (Hill et al. 1993) and passively floating objects therein, the backward simulations were
252 performed individually for each cluster. The influence of the wind drag factor on the origin of
253 items from the different source regions was analysed separately for the nearshore and the
254 offshore cluster using χ^2 -tests based on 3 (number of wind drift factors) \times 6 (number of source
255 regions) contingency tables.

256

257 *Factors explaining seafloor litter distribution*

258 To understand the processes that shape the distribution of litter on the seafloor we
259 investigated six environmental and anthropogenic factors for their role as sources of benthic
260 marine litter and as drivers of near-bottom dispersal: bottom trawling intensity, distance to
261 major shipping routes, bottom shear stress, mean horizontal surface velocities, water depth,
262 and distance to coast. Data on bottom trawling intensity were available for the year 2006 from
263 vessel monitoring system data (Schröder et al. 2008). For our analysis we assume that the data
264 from 2006 are representative for the long-term distribution and intensity of the bottom
265 trawling activity in the south-eastern North Sea. Distances to shipping routes were calculated
266 from long-term annual shipping densities collected as AIS data by Baldock (2009). Bottom
267 shear stress can lead to the resuspension and near-bottom transport of deposited material, such
268 as sediment or litter objects. It is a composite vectorial force resulting from the joint action of
269 vertical wave-induced lift and horizontal transport by tidal currents. The composite bottom
270 shear stress was calculated from the specific values for shear stress induced by wave action
271 and by bottom currents of the year 2007, assuming again long-term validity of the data. The
272 specific shear stress values were retrieved from the coastMap Geoportal (www.coastmap.org)
273 under CC BY-NC 4.0 license. For each day, only the higher one of the two shear stress values
274 was used, assuming that the movement of a particle or object near the seafloor is primarily
275 determined by the dominant force. From the daily values the 0.95 quantile was chosen as
276 bottom shear stress characteristic. The mean horizontal surface velocities were calculated as
277 an average of the velocity dataset used for the above simulations of trajectories of Lagrangian
278 particles. Water depth was taken from GEBCO bathymetry (Wetherall 2015). Distance from
279 coast is a simple derivative of a shoreline map. The explanatory variables were processed

280 using R v.3.0.3 (R Core Team 2014) with a spatial resolution of 1×1 km.

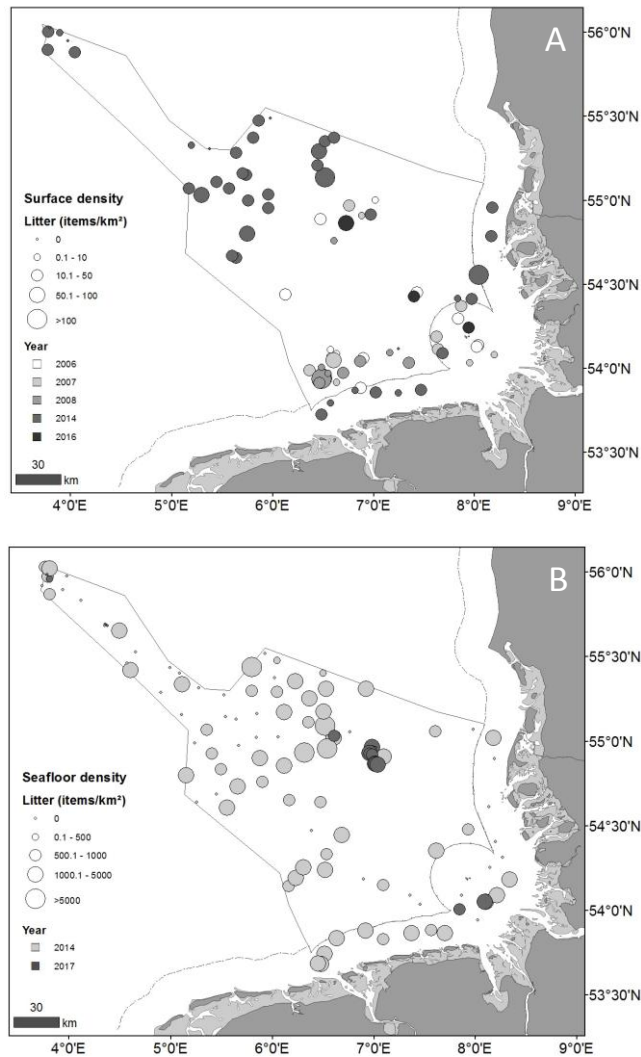
281 We used two different methods to investigate the potential of the selected variables to
282 explain the distribution of seafloor litter: Maximum entropy (Maxent) (Phillips et al. 2006)
283 and Random Forest (RF) (Breiman 2001). Maxent was applied to presence data. RF was
284 applied to presence/absence data and density data to study the quantitative relationship
285 between the spatial distribution of litter and the environmental variables. Model accuracy was
286 assessed as the Receiver Operating Characteristic Area Under Curve (ROC AUC), Cohens
287 Kappa (Kappa), Brier score and root mean squared error (RMSE). We used null models to test
288 if data locations were spatially biased with regard to the predictors. Here, performance metrics
289 of models of randomized data act as a baseline to identify the metrics a null hypothesis
290 situation might have. Non-trivial patterns must then significantly deviate from null model
291 performance.

292

293 **Results**

294 *Distribution of floating litter*

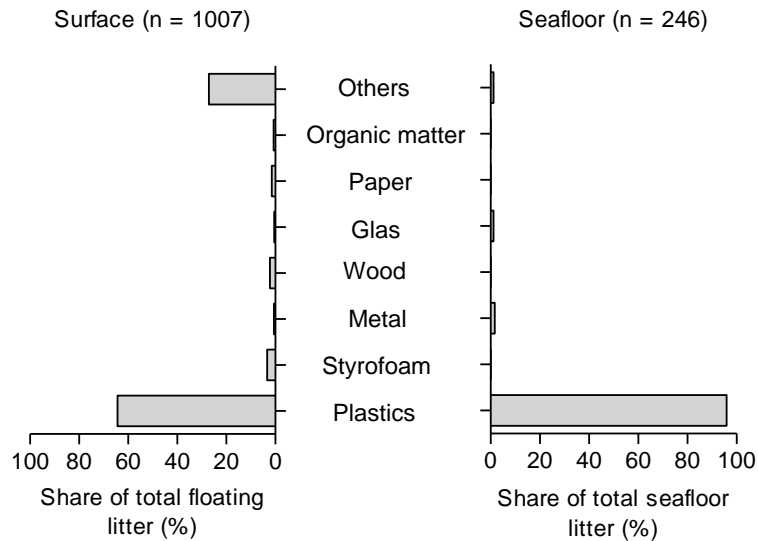
295 Floating marine litter occurred along the entire German coast as well as in offshore
296 waters (Figure 2A). Floating litter was observed on 74 out of 78 surface transects. Densities of
297 floating litter ranged from 1 to 272 items km^{-2} (mean \pm SD = 30.6 ± 41.8 items km^{-2} ; median:
298 19.7 items km^{-2}) with highest densities occurring in the North Frisian coastal waters, in the
299 area of the Borkum Reef Ground in the south-western part of the German Exclusive
300 Economic Zone (EEZ), and in the central-northern part of the German EEZ off the coast of
301 southern Denmark. Data from the southern and south-western coastal waters of the German
302 North Sea, where floating litter was surveyed repeatedly over several years, indicate
303 substantial temporal variability in litter densities. No clear gradient in litter densities from
304 coastal to offshore waters was evident.



305

306 **Figure 2** Densities of (A) floating and (B) benthic litter in the south-eastern North Sea.

307 Floating litter was dominated by plastics (Figure 3) which held a share of 64% of the
 308 total amount of litter. Floating plastics consisted of numerous small plastic pieces and
 309 fragments but also of larger items, such as plastic bottles and plastic bags. The next most
 310 common categories of floating litter were styrofoam, wood and paper. Only few items
 311 consisted of glass, metal or organic matter. The relatively large fraction of “Other” floating
 312 litter objects mostly comprised unidentifiable floating objects but also cigarette butts.



313

314 **Figure 3** Composition of floating (left-hand side) and benthic (right-hand side) litter in the
 315 south-eastern North Sea. The litter composition differed significantly between the surface and
 316 the seafloor ($df = 6$; $\chi^2 = 65.1$; $p < 0.01$).

317

318 *Distribution of seafloor litter*

319 Seafloor litter was found in both coastal and in offshore waters of the German North
 320 Sea (Figure 2B). Marine litter was found in 70 out of 122 bottom trawls. Densities of litter at
 321 the seafloor varied between 96 and 33,675 items km^{-2} (mean \pm SD = 1307.3 ± 3352.3 items
 322 km^{-2} , median: 610 items km^{-2}) with highest densities occurring in the central-northern part of
 323 the German EEZ off the west coast of southern Denmark. Seafloor litter was found in coastal
 324 waters along the East Frisian coast and off the mouths of the rivers Elbe and Weser but was
 325 almost entirely absent along the North Frisian coast. Similarly, in the north-western stretches
 326 of the German EEZ a considerable number of stations were free of seafloor litter.

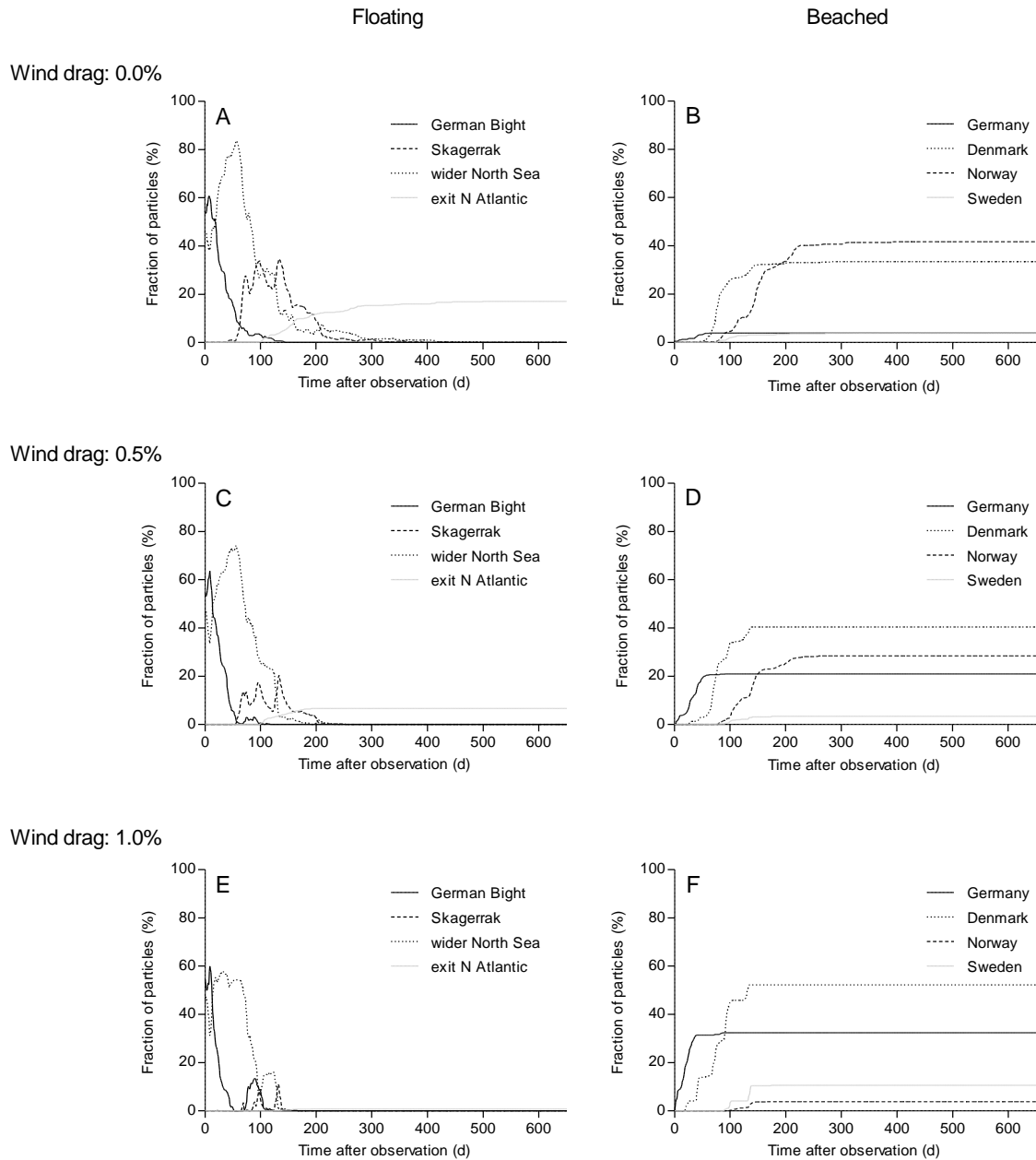
327 Litter on the seafloor consisted almost exclusively of plastics (Figure 3). The by far
 328 most common plastic items on the sea floor were fibers from fishing nets but pieces of plastic
 329 foil and other packaging material made of plastics were also common. Overall, fishing related
 330 litter items made up 76% of the total seafloor litter. Few metal and glass items were
 331 encountered while paper, wood, styrofoam and organic material were entirely absent from the
 332 seafloor samples. The few “Other” litter items were pieces of coal slag and a tire (consisting
 333 of rubber and metal). Although plastics dominated the litter at both the surface and the
 334 seafloor the overall litter composition differed significantly between the two habitats ($df = 7$;
 335 $\chi^2 = 111.7$; $p < 0.01$). Moreover, densities of litter at the surface and at the seafloor were not
 336 correlated with each other (Spearman rank $r = 0.26$; $p = 0.23$).

337

338 *Drift trajectories of floating litter*

339 The simulation showed that the trajectories of the floating particles in the south-
340 eastern North Sea are typically characterized by an initial northward drift resulting in a rapid
341 initial drop of particles in the German Bight and an increase in particle densities in the wider
342 North Sea, i.e. the area of the North Sea outside the German Bight, and in the Skagerrak
343 (Figure 4). On their way north a considerable number of particles was deposited along the
344 Danish west coast or forced into the Skagerrak where they were trapped by eddies between
345 the incoming North Sea water and the outflow from the Baltic. Consequently, the fraction of
346 particles floating in the Skagerrak increased while the proportion of particles in the wider
347 North Sea decreased. In the Skagerrak particles may be deposited on the Danish, the
348 Norwegian or the Swedish coast. When released from the Skagerrak the particles drifted north
349 to become either deposited along the Norwegian coast or to exit the North Sea into the North
350 Atlantic.

351 The drift trajectory and, thus, the probability of being deposited on a shore or
352 transported out of the North Sea into the North Atlantic strongly varied with the wind drag
353 coefficient (Figure S2). Particles which were transported solely by currents (wind drag =
354 0.0%) had a relatively high chance of entering the North Atlantic (Figure 4A) or being
355 deposited along the Norwegian North Sea coast (Figure 4B). Particles which were more
356 strongly controlled by wind were more rapidly pushed eastward towards the coast by the
357 prevailing westerly winds (Figure 4C-F). Consequently, at a wind drag coefficient of 1.0% a
358 considerable number of particles did not leave the German Bight but became quickly washed
359 ashore along the German and the Danish coast. The probability of particles beaching at the
360 Swedish coast also increased with the wind drag coefficient whereas particles rarely reached
361 the Norwegian coast or left the North Sea when a high wind drag coefficient was used. The
362 fraction of particles deposited on the Danish coast was generally high at about 40% of the
363 total number of particles floating in the German North Sea.

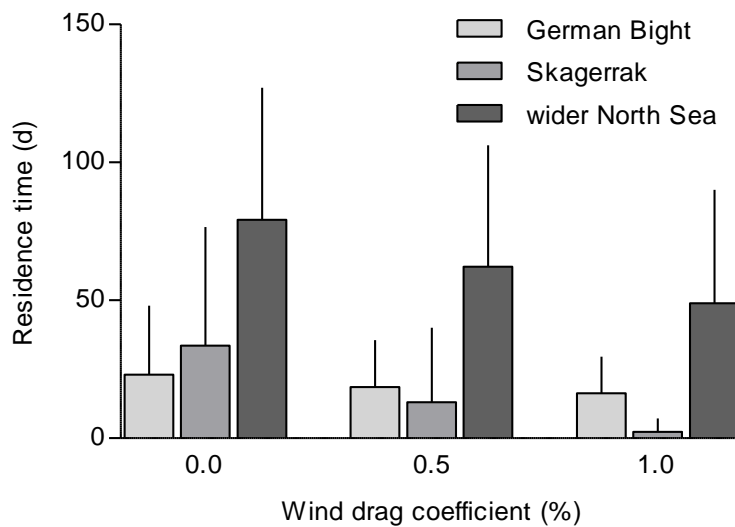


364

365 **Figure 4** Results of the drift simulation of Lagrangian particles in the North Sea. Simulations
 366 were run for particles using different wind drag coefficients (A, B: 0.0%; C, D: 0.5%; E, F:
 367 1.0%). A, D and E show the temporal variations in the number of active particles floating in
 368 different regions of the North Sea. B, D and F show the number of particles beached in
 369 different coastal regions of the North Sea. Each particle represents a specific floating litter
 370 item observed during the survey and for each particle the simulation starts at the time when it
 371 was observed (day 0). For each day of the simulation the proportions of floating and beached
 372 particles, respectively, sum up to 100%.
 373

374 The residence time of floating items was longest in the wider North Sea. The average
 375 residence time of floating items in the German Bight was largely independent of the wind
 376 drag coefficient and varied between 16 and 23 days (Figure 5). In both the Skagerrak and the

377 wider North Sea the average residence time of floating items declined continuously with
378 increasing wind drag factor from about 33 to 2 days and from 79 to 48 days, respectively.



379

380 **Figure 5** Results of the drift simulation of Lagrangian particles in the North Sea: average
381 residence times of particles of different wind drag coefficients in different regions of the
382 North Sea. Error bars indicate standard deviation.
383

384 The backward simulation revealed that particles from the offshore and the nearshore
385 cluster had different source regions (Figure S3). Depending on the wind drag coefficient, 87-
386 96% of the particles from the nearshore cluster originated from the German and Dutch coast,
387 with major contributions from the rivers Ems, Jade, Weser and Elbe, whereas only 14-27% of
388 the particles from the offshore cluster came from the German and the Dutch coast (Table 1).
389 The majority of particles from the offshore cluster originated from more distant western
390 source regions with major contributions from the British Channel and the river Rhine but also
391 from the coasts of France and the British Isles. Generally, only a small number of particles
392 originated from the relatively short Belgian coastline.

393 The wind drag coefficient influenced the contribution of the different source regions,
394 particularly the contributions from the German and the Dutch coasts (Figure S4). The
395 proportion of particles originating from the German coast decreased with increasing wind
396 drag factor whereas the Dutch coastal zone had a higher contribution of items with elevated
397 wind drag factor. For the contribution from the more distant source regions the effect of the
398 wind drag factor was less pronounced. The effect of the wind drag factor was statistically
399 significant for both the nearshore ($\chi^2 = 90.3$, $df = 10$, $p < 0.01$) and the offshore cluster ($\chi^2 =$
400 54.6 , $df = 10$, $p < 0.01$).

401

402 **Table 1** Results of the backward drift simulation of Lagrangian particles in the North Sea:
 403 source regions of particles from the nearshore and the offshore cluster computed using
 404 different wind drag coefficients. The numbers are percentage values according to the share of
 405 particles originating from the respective source region.

Wind drag coefficient	Nearshore (n = 507 items)			Offshore (n = 265 items)		
	0.0%	0.5%	1.0%	0.0%	0.5%	1.0%
France	0.39	2.17	0.59	25.28	30.94	16.60
British Isles	0.00	8.88	7.69	30.94	20.76	34.34
Belgium	0.20	0.20	0.00	0.00	1.89	0.38
The Netherlands	36.89	33.53	44.58	15.85	15.09	22.64
Germany	60.55	54.83	46.75	9.06	7.92	3.40
Floating	1.97	0.39	0.39	18.87	23.40	22.64

407

408 *Factors explaining seafloor litter distribution*

409 The models used to identify factors that shape the distribution of seafloor litter showed
 410 only moderate quality in terms of ROC AUC and Kappa (Table 2). Both Brier scores for the
 411 probabilistic models from presence-only and presence/absence data and RMSE for the density
 412 predictions suggest rather low explanatory power. According to the sensitivity analysis based
 413 on Jansen estimators (Jansen 1999) shear stress was the most influential variable whereas
 414 distance to major shipping routes was less important.

415 Null models calculated from 5-fold cross validation runs revealed no spatial bias with
 416 respect to the explanatory variables employed. Spatial autocorrelation was low. For a lag
 417 distance of 30 nm the Moran I value was 0.25 ($p < 0.001$). This can be attributed to the
 418 predictors used, as the Moran I test of the residuals was negative (0.06 at $p = 0.21$).

419

420 **Table 2** Diagnostic model accuracy metrics (\pm SD) for litter at seafloor form repeated 5-fold
 421 cross validation.

Response	Method	ROC AUC	Kappa	RMSE	Brier score
Presence	Maxent	0.48 \pm 0.06	0.18 \pm 0.08	-	0.31 \pm 0.00
Presence/absence	Random forest	0.55 \pm 0.13	0.27 \pm 0.15	-	0.31 \pm 0.07
Density	Random forest	0.57 \pm 0.13	0.29 \pm 0.15	1590 \pm 73	-

422

423

424

425 **Discussion**

426 The results of this study reveal considerable quantities of marine litter at the sea
 427 surface and on the seafloor of the south-eastern North Sea. The great majority of marine litter

428 consisted of plastics, and fisheries activities were identified as a major source of marine litter.
429 The drift simulations indicated that a large fraction of the floating marine litter in the North
430 Sea is deposited on the beaches of Germany and Scandinavian countries. However, the factors
431 which control the distribution of litter on the seafloor are still poorly understood. The
432 identification of major sources and sites of deposition of marine litter may help developing
433 strategies to reduce the pollution of the North Sea.

434

435 *Composition and distribution of floating and benthic marine debris*

436 It is well established since decades that the North Sea is substantially polluted with
437 marine anthropogenic litter (Vauk & Schrey 1987). Since the first survey by Dixon and Dixon
438 (1983) the density of floating litter in the North Sea has increased by an order of magnitude to
439 an average of about 30 items km⁻² (Thiel et al. 2011). The data from the recent survey by
440 Thiel et al. (2011) have been incorporated in our dataset to achieve a temporally more
441 integrated estimate of the distribution and composition of floating litter with a higher spatial
442 coverage, which can be contrasted with the densities on the seafloor.

443 The average litter density on the seafloor was about 40 times higher than at the sea
444 surface. Direct comparisons of surface and seafloor densities of marine litter are still scarce.
445 Globally, however, litter densities seem to be generally higher in benthic than in pelagic
446 environments (Galgani et al. 2015). In our study the difference may have been amplified by
447 the different methods applied to quantify litter in the two habitats. To enhance the
448 comparability of the densities, towed devices with similar mesh size could have been used to
449 quantify litter at the surface (neuston net) and on the seafloor (beam trawl). The use of a
450 neuston net may have led to higher surface densities. However, we would expect only a minor
451 effect on the overall distribution pattern of floating litter because the majority of small items
452 collected from the seafloor originated from bottom trawling activities and were directly
453 deposited in the benthic environment. Moreover, due to the relatively small net aperture litter
454 densities estimated with a neuston net are sensitive to spatial clumping of floating litter (Ryan
455 et al. 2009). Therefore, visual ship based surveys are a common, non-invasive method for the
456 quantification of floating macrolitter (Thiel & Gutow 2005).

457 Higher litter densities on the seafloor than at the surface indicate (a) continuous
458 vertical import of marine litter from the sea surface and the water column and/or (b) direct
459 deposition of litter on the seafloor. The average density of about 1300 items km⁻² on the
460 seafloor was about one to two orders of magnitude higher than average seafloor densities

461 reported in previous studies from the North Sea including the German EEZ (Galgani et al.
462 2000: 156 items km⁻², Kammann et al. 2017: 11-24 items km⁻², Maes et al. 2018: 40-49 items
463 items km⁻²). The previous studies presented data from fishery surveys (e.g. ICES international
464 bottom trawl survey – IBTS) which uses nets with mesh sizes of 20-40 mm in the cod-end
465 and, hence, the authors suggested that the amount of litter on the seafloor was probably
466 underestimated (Galgani et al. 2000, Kammann et al. 2017). The beam-trawl with a mesh size
467 of 10 mm used in our study probably produced a better estimate of macrolitter quantities on
468 the seafloor of the North Sea, which was in the range of some densities measured in heavily
469 polluted regions such as the South China Sea (Zhou et al. 2011) and the Mediterranean
470 (Tubau et al. 2015).

471 In both habitats, the surface and the seafloor, the majority of litter items consisted of
472 plastics. The proportion of 64% at the surface is lower than the global contribution of plastics
473 to the overall amount of litter in the world ocean of 73% (Bergmann et al. 2017a). The share
474 of plastics of 95% on the seafloor is clearly above previous estimates of a 50-80%
475 contribution of plastics to the benthic litter in the North Sea (Galgani et al. 2000, Schulz et al.
476 2015a, Kammann et al. 2017, Maes et al. 2018). The composition of litter differed
477 significantly between the surface and the seafloor. Furthermore, litter densities did not
478 correlate among the two habitats indicating that the contribution of vertical import from the
479 surface to the benthic litter may be negligible relative to the direct deposition in the benthic
480 habitat. Additionally, near-bottom transport processes may have redistributed benthic items
481 thereby obscuring the spatial pattern of litter deposition on the seafloor.

482 The majority of litter items on the seafloor were related to fishing activities including
483 numerous fibers and yarn from fishing nets and dolly ropes clearly pointing at bottom
484 trawling as the major source of benthic litter. Fisheries have previously been identified as an
485 important source of marine litter on the seafloor and on beaches of the North Sea (Galgani et
486 al. 2000, Schulz et al. 2015a) and in other regions of the world ocean (Merrell 1984, Walker
487 et al. 1997, Edyvane et al. 2004, Buhl-Mortensen & Buhl-Mortensen 2017). An analysis of
488 data from OSPAR beach monitoring indicated that the amount of fishery related litter on the
489 coasts of the North-East Atlantic, including the German North Sea, is decreasing presumably
490 due to declining coastal fishery activities (Schulz et al. 2013, 2015b). However, our data
491 reveal that the contribution of fishery related items to the overall benthic litter in the south-
492 eastern North Sea has increased from 60% (Schulz et al. 2015a) to 76% (this study). Schulz et
493 al. (2015a) reported litter quantities by weight. Therefore, the comparability of the share of
494 fishery related debris among the two studies is limited. Bergmann et al. (2017b) suggest that

495 quantifying marine debris by weight may overestimate the contribution of fisheries activities
496 because of the potential occurrence of exceptionally heavy fishing gear or remains thereof.
497 Accordingly, the increasing share of fishery related debris may indicate a disproportional and
498 rapid accumulation of this litter category in the North Sea probably due to chronically high
499 bottom trawling activity in this region (Stelzenmüller et al. 2014, Kenny et al. 2017).

500

501 *Drift trajectories of floating litter*

502 It is impossible to determine where exactly an object, which was observed floating at
503 the sea surface, has entered the marine environment. Hence, each floating object observed
504 during our survey could have started its floating journey at any point along the modelled
505 backward trajectory of the corresponding Lagrangian particle. Nevertheless, the simulation
506 permitted to identify two clusters of particles which differ considerably in their backward
507 trajectories, pointing at different source regions. The offshore cluster had a larger proportion
508 of particles that could be traced back to more distant western source regions at the British
509 Isles, France and the English Channel whereas the great majority of particles in the nearshore
510 cluster could be traced back to nearby source regions in the Netherlands and Germany.
511 Accordingly, the distribution of floating marine debris in the south-eastern North Sea is
512 decisively determined by the geographic location of the sources (see also Hainbucher et al.
513 1987). A large proportion of the particles were discharged from rivers into the North Sea
514 confirming that riverine transport substantially contributes to the littering of coastal seas
515 (Rech et al. 2014). The rivers Ems, Jade, Weser and Elbe along the Dutch and German coast
516 are major sources of particles of the nearshore cluster. The offshore cluster is fed by the Rhine
517 and the East Anglia Plume, which is responsible for an eastward transport of particles
518 discharged by the rivers Thames and Humber into the central North Sea (supplementary
519 material S5, see also Pietrzak et al. 2011).

520 The two clusters were located on different sites of the major salinity front of the
521 German Bight (Figure S1). Previous studies have shown that oceanic features, such as fronts
522 and convergences, can efficiently constrain the dispersal and mixing of flotsam. For example,
523 large amounts of marine litter were observed below a salinity front in the Rio de la Plata
524 estuary (Argentina) indicating that floating items are retained by the front before they sink to
525 the bottom (Acha et al. 2003). Similarly, floating kelps have been found to accumulate along
526 an estuarine front in the Chilean fjord system (Hinojosa et al. 2010). Likely, the distribution
527 pattern of floating marine debris in the south-eastern North Sea, which is determined by the

528 origin of the flotsam, is consolidated by the frontal system that may prevent mixing of the two
529 clusters.

530 The further trajectories of particles in the German Bight are strongly influenced by
531 wind drag. The anticlockwise current system of the North Sea transports particles northward
532 out of the German Bight (Hainbucher et al. 1987, Schönfeld 1995) while prevailing westerly
533 winds push particles towards the shore (Neumann et al. 2014) suggesting an early beaching of
534 buoyant objects with considerable freeboard, such as buoys and PET drink bottles. Contrarily,
535 particles with a low wind drag coefficient, analogous to objects which do not extend above
536 the surface such as plastic bags and foils, are less strongly forced by wind and become later
537 deposited on the shore or are transported by currents out of the North Sea into the North
538 Atlantic. Hence, differential transport likely results in a segregation of objects with different
539 floating behavior (Laxague et al. 2018). The interaction of currents and wind may lead to a
540 fractionation of deposited items along the eastern shore of the German Bight, with objects
541 with high freeboard in the south and objects with less or no freeboard further north. Denmark
542 is situated in the centre of this coastal deposition line and receives a high share (about 40%) of
543 particles irrespective of the wind drag coefficient. Consequently, the western coast of
544 Denmark appears to be a major sink region for floating marine debris in the German Bight.
545 However, the deposition of particles on the shore is a simplifying assumption as the actual
546 behavior of floating objects in complex and turbulent coastal currents cannot be modelled yet.
547 Moreover, the final deposition pattern along the coast is probably influenced by resuspension
548 of stranded items which is not considered in our drift trajectory model. However, Thiel et al.
549 (2003) observed that a large fraction of the litter found in coastal waters in the SE Pacific was
550 of local origin indicating that objects floating in coastal waters unlikely escape to offshore
551 waters but have a high probability of being washed on the nearby shore.

552 The trajectory and the destination of a floating object strongly depend on its buoyancy
553 and persistence at the sea surface (Thiel & Gutow 2005). Loss of buoyancy and sinking of
554 floating litter can be accelerated by biofouling when colonizing organisms enhance the
555 specific gravity of an object (Ye & Andrady 1991) especially in objects with a high
556 surface:volume ratio (Chubarenko et al. 2016). Submerged objects are more densely
557 colonized by fouling organisms than objects with large freeboard that substantially extend
558 above the sea surface (Bravo et al. 2011). Accordingly, objects floating below the surface,
559 such as plastic bags, are probably at particular risk of losing buoyancy. Similar to the results
560 of Schönfeld (1995) our drift simulation showed that many particles with low wind drag
561 coefficient are transported out of the German Bight and trapped for a considerable period of

562 time in the Skagerrak. The extended residence time in the Skagerrak likely enhances the
563 probability of local sinking and deposition on the seafloor. Already in the 1970s, Holmström
564 (1975) reported on abundant plastic bags and foils on the seafloor of the Skagerrak. Typically,
565 those plastic sheets showed considerable colonization by sessile organisms which have
566 probably induced sinking. Accordingly, the Skagerrak and particularly its seafloor seem to
567 represent another important sink for floating marine litter from the south-eastern North Sea.

568 Particles, which are released from the Skagerrak, are mostly transported northward
569 along the Norwegian coast to exit the North Sea into the North Atlantic. Huge amounts of
570 floating marine litter entering the North Atlantic in European waters are travelling towards the
571 Arctic (Cózar et al. 2017). Depending on buoyancy, persistence and wind drag of the objects,
572 floating marine litter from the German Bight may substantially contribute to the pollution of
573 the sensitive Arctic marine ecosystem.

574

575 *Factors explaining seafloor litter distribution*

576 Our models did not confirm a correlation of the distribution of marine litter on the
577 seafloor with any of the selected variables. Accordingly, a clear identification of the sources
578 of benthic litter based on these results was not possible although the composition of the
579 seafloor litter collected during our survey clearly hint at fisheries as a major source. This
580 suggests that the distribution of seafloor litter in the south-eastern North Sea is rather
581 governed by near-bottom transport processes rather than by the source regime and is
582 corroborated by our finding that bottom shear stress influences the distribution of benthic
583 litter. Schulz et al. (2015a) also suggested near-bottom transport to be an important process
584 shaping the distribution of litter on beaches and on the seafloor of the south-eastern North
585 Sea. Considerable near-bottom transport is also indicated by the distribution of benthic litter
586 on larger spatial scales. Pham et al. (2014) observed particularly high densities of benthic
587 litter in submarine canyons indicating re-mobilization of benthic litter by near-bottom currents
588 and movements down the slope of the seafloor.

589

590 *Conclusions*

591 Our survey together with the results from model simulations allowed for identifying
592 important sources, distribution and trajectories of marine litter in the south-eastern North Sea.
593 Similar to findings from previous studies we identified fisheries as a major source especially

594 of seafloor litter. Therefore, reducing the input of litter from fisheries would substantially
595 reduce the rapid accumulation of litter in the North Sea environment.

596 The distribution of floating litter in the North Sea is primarily determined by the origin
597 of the litter whereas the destination and final deposition is strongly dependent on the wind
598 drag coefficient of the litter items. Major rivers are releasing substantial amounts of floating
599 litter into the south-eastern North Sea whereas the west coast of Denmark as well as the
600 Skagerrak could be identified as major sink regions. Frequent clean-ups of the Danish
601 coastline would allow for the removal of substantial amounts of litter from the marine
602 environment. There is common agreement that the unselective extraction of litter with heavy
603 gear can induce substantial damage to sensitive marine ecosystems. However, the negative
604 environmental implications may be minimized by the strategic and careful removal of
605 stranded litter from easily accessible coastal locations where litter is known to accumulate.

606

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616

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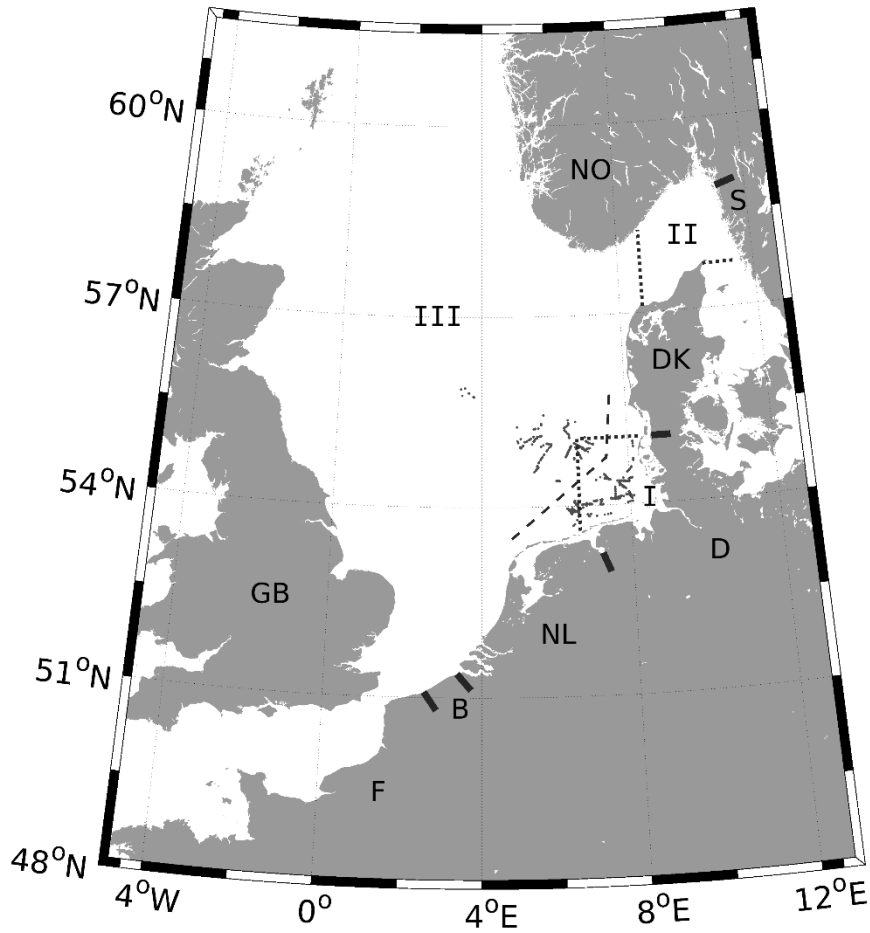
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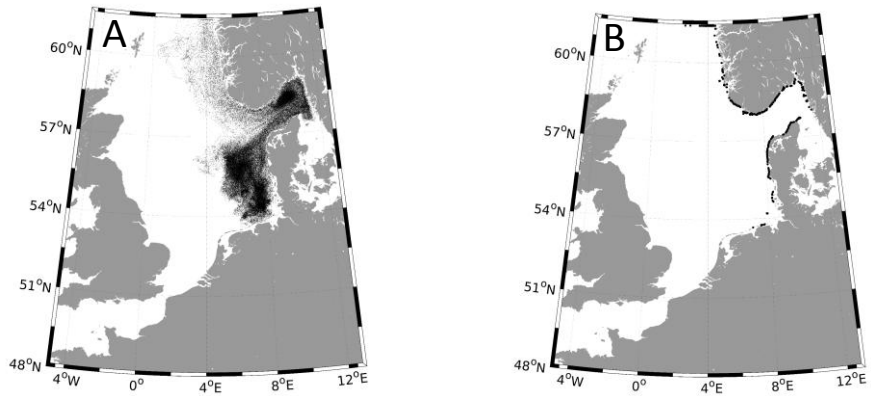
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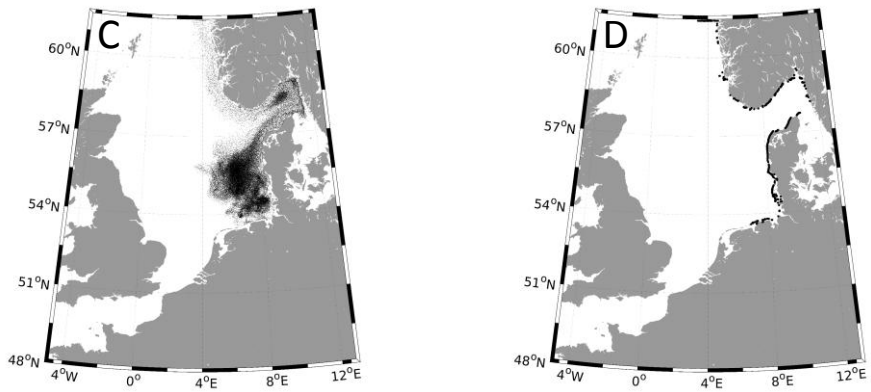


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 828 **Figure S1** Map of the North Sea. Small dots in the south-eastern North Sea denote the
 829 positions where floating litter items were observed during the survey. Floating objects of the
 830 nearshore cluster and the offshore cluster are separated by the major salinity front of the
 831 German Bight (dashed line). Borders between marine regions (I: German Bight, II: Skagerrak,
 832 III: wider North Sea) are indicated as dotted lines. National frontiers between coastlines of
 833 different countries are indicated as bold black marks.
 834

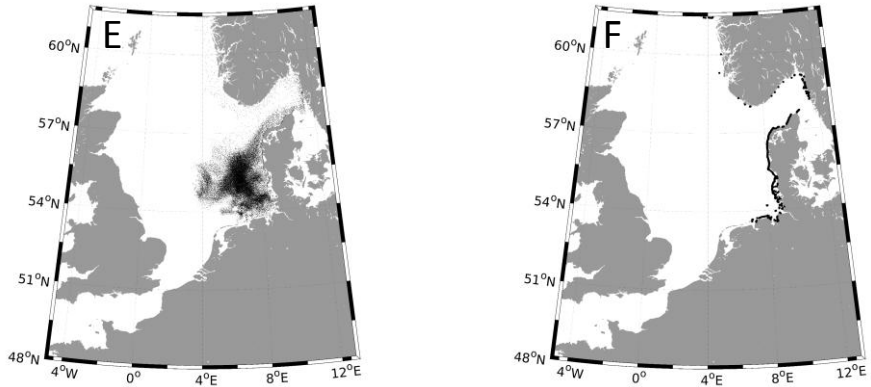
Wind drag coefficient: 0.0%



Wind drag coefficient: 0.5%



Wind drag coefficient: 1.0%



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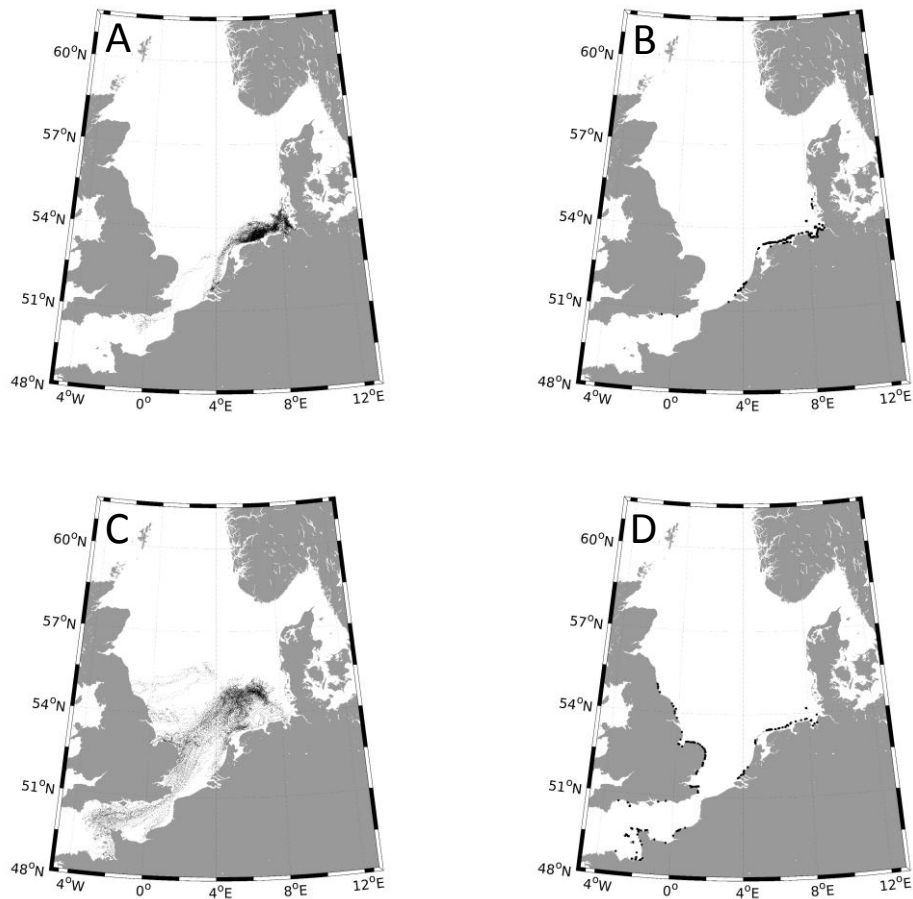
836 **Figure S2** Results of the forward drift simulation of all 772 Lagrangian particles in the North
837 Sea: trajectories (A,C,E) and sites of deposition (B,D,F) of particles using different wind drift
838 coefficients (A,B: 0.0%; C,D: 0.5%; E,F: 1.0%)

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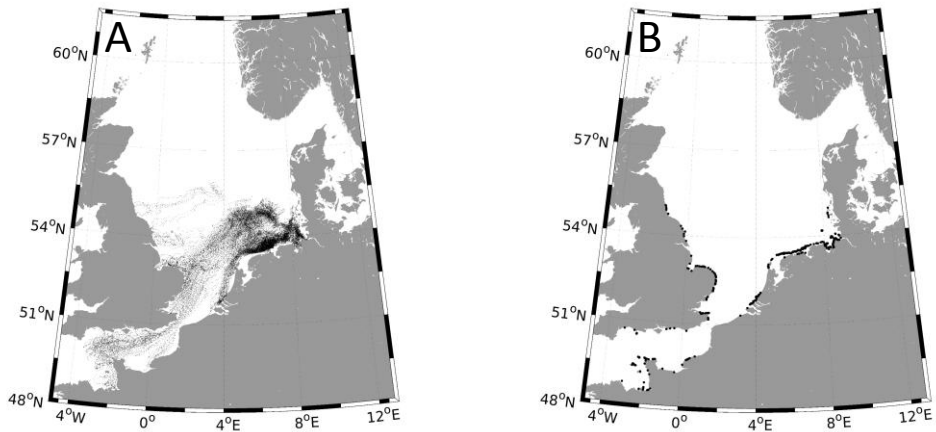
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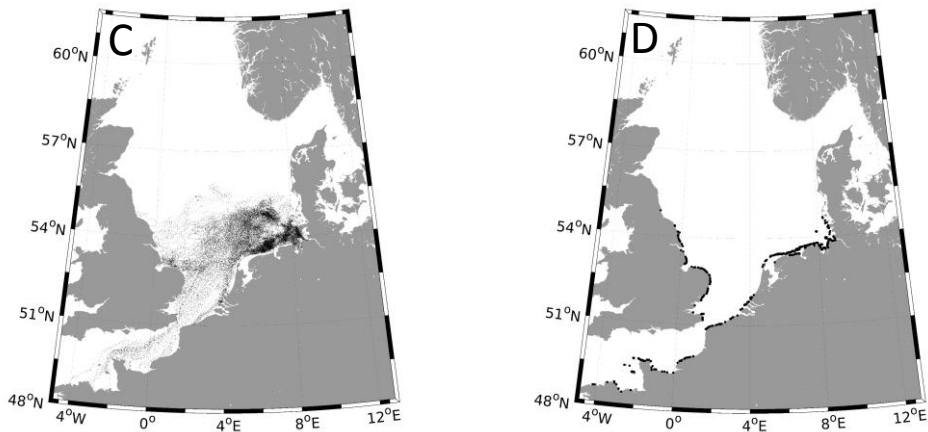
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844 **Figure S3** Results of the backward drift simulations of Lagrangian particles in the North Sea:
 845 A) backward trajectories for particles observed in the nearshore cluster, B) backward
 846 trajectories for particles observed in the offshore cluster, C) computed source (or starting)
 847 point for the particles later observed in the nearshore cluster and D) same as in C) but for the
 848 offshore cluster. Both simulations used a wind drag coefficient of 0.0%.

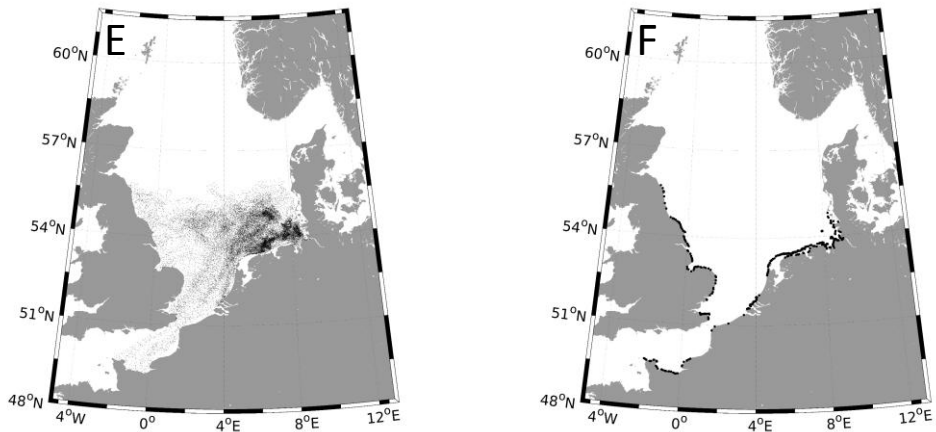
Wind drag coefficient: 0.0%



Wind drag coefficient: 0.5%



Wind drag coefficient: 1.0%



849

850 **Figure S4** Results of the backward drift simulations of all 772 Lagrangian particles in the
851 North Sea for different wind drag coefficients: (A,C,E) backward trajectories and (B,D,F)
852 computed source (or starting) point of the particles representing observed floating litter items.
853 Wind drag coefficients in those simulations are 0.0% (A,B), 0,5% (C,D) and 1.0% (E,F).

854 **Animation S5** Drift animation of Lagrangian particles discharged by major rivers into the
855 southern North Sea. Particles are released at intervals of 12 h and wind drag was set to 0.0%.
856 The animation shows the particle positions every 24 h over a time period of 1 year. The
857 velocity data input was the same as for the litter simulations. Beached particles are not
858 resuspended but stay at the positions where they enter land.